

Environmental Progress  
& Sustainable Energy

**CARBON FOOTPRINT AND ENERGY ANALYSIS OF BIO-CH<sub>4</sub>  
FROM A MIXTURE OF FOOD WASTE AND DAIRY MANURE IN  
DENVER, COLORADO**

Journal:	<i>Environmental Progress</i>
Manuscript ID	Draft
Wiley - Manuscript type:	Original Manuscript
Date Submitted by the Author:	n/a
Complete List of Authors:	Ankathi, Sharath; Michigan Technological University, Chemical Engineering Shonnard, David; Michigan Technological University, Chemical Engineering; Michigan Technological University, Sustainable Futures Institute Potter, James; AG Energy USA, LLC
Keywords:	Life Cycle Assessment, Waste Management, Biogas, Anaerobic
Alternate Keywords:	

SCHOLARONE™  
Manuscripts

# CARBON FOOTPRINT AND ENERGY ANALYSIS OF BIO-CH<sub>4</sub> FROM A MIXTURE OF FOOD WASTE AND DAIRY MANURE IN DENVER, COLORADO

Sharath K Ankathi<sup>1</sup>, James S Potter<sup>2</sup>, David R Shonnard<sup>1,3</sup>

*1 Department of Chemical Engineering, Michigan Technological University, Houghton, MI*

*2 AG Energy USA, LLC, Hampton, NH*

*3 Sustainable Futures Institute, Michigan Technological University, Houghton, MI*

## Abstract

Anaerobic digestion (AD) is a possible alternative to landfilling of food waste and conventional manure management in order to reduce methane emissions. Key results of this carbon footprint show that the AD Bio-CH<sub>4</sub> pathway has 15.5% lower greenhouse gas (GHG) emissions compared to the prior practice of composting of food waste and manure in Denver, CO. Results from modeling the GHG emissions for Bio-CH<sub>4</sub> production from AD conversion of food waste and manure with avoiding of food waste landfilling and conventional management of dairy manure emits -3.5 kg CO<sub>2</sub> equivalents / kg Bio-CH<sub>4</sub> assuming the electricity was generated using collected landfill gas. This emission intensity is favorable compared to that of fossil natural gas of 4.3 kg CO<sub>2</sub> equivalents / kg Bio-CH<sub>4</sub> equivalents of fossil natural gas. Effects of life cycle system parameters on GHG results are investigated in scenario analysis as well as time dependent analysis of avoided landfill emissions.

## Novelty or Significance

A novel feature of this Bio-CH<sub>4</sub> LCA that distinguish it from previous studies are that input data for the anaerobic digestion (AD) process is from an actual operating facility; Heartland Biogas LLC in LaSalle, CO, rather from either modeled data or smaller scale processing. A second novel aspect is a transient analysis on avoided landfill emissions when food waste is diverted to the Bio-CH<sub>4</sub> process, whereas prior work used steady-state analysis. The LCA includes the effects of an innovative water

1  
2 recycle scenario where digestate from the AD process containing mineral nutrients is applied to local  
3  
4 crop lands supporting dairy operations.  
5

### 6 7 **Keywords**

8  
9 Keywords: life cycle assessment, carbon footprint, anaerobic digestion of food waste and  
10  
11 manure, transient analysis  
12

### 13 14 **Introduction**

#### 15 16 ***Food waste generation***

17  
18 The majority of municipal solid waste (MSW) is from the industrial, commercial and residential  
19  
20 sectors, which together account for 254 million short tons per year according to recent statistics in the  
21  
22 US <sup>1</sup>. Organic materials constitute the majority portion of MSW in which 27% is paper and paperboard  
23  
24 with food waste and yard trimmings accounting for 28%. The remainder is made up of plastic > metal >  
25  
26 rubber-leather-textiles > wood > glass. The generated MSW is managed by different methods out of  
27  
28 which only 12% is processed through incineration and energy collection systems, 53% is disposed to  
29  
30 landfills, and 35% goes to recycling and composting <sup>2</sup>. In the USA, out of all the generated MSW the  
31  
32 second highest component is organic food waste with 14.6% of total waste. When disposed in a landfill,  
33  
34 food waste releases the highest amount of methane emissions per dry weight of disposed materials.  
35  
36 Furthermore, the wastage of food is about 30-40% of the food supply, equaling more than 20 pounds of  
37  
38 food per person per month. With the conversions of dry solids in food waste to methane, 216 Mg of dry  
39  
40 food waste generates 4700 dekatherms of Bio-CH<sub>4</sub> through anaerobic digestion (AD), which is  
41  
42 equivalent to 105.19 Mg of fossil natural gas (by taking the lower heating value of 47.14 MJ/kg).  
43  
44 According to 2013 EPA report the US generates 36 million wet tons / yr of food waste, which if entirely  
45  
46 converted to Bio-CH<sub>4</sub> has a potential to replace .74% of US total natural gas usage of 548 million tons  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

1  
2 annually (see Supporting information section-4 for calculations). In the US, 97% of the total food waste  
3  
4 is buried in landfills where it causes odor as decomposes and produces methane <sup>3</sup>.

### 6 7 ***Landfills and GHG emissions***

8  
9 Currently in the US, 44% of the landfills use gas collection systems plus flaring to reduce GHG  
10  
11 emissions and there are about 850 such flaring landfills out of a total 1908 landfills <sup>4</sup>. The landfill  
12  
13 methane outreach program (LMOP) currently tracks new landfill gas to energy projects in the US, which  
14  
15 reports that there are 652 such landfills turning the methane into useful energy sources (electricity  
16  
17 mostly). From 2014 statistics, total methane emissions from landfills in the US are 102.8 million metric  
18  
19 tons CO<sub>2</sub> eq. <sup>5</sup>. Methane is more efficient in trapping infrared radiation than CO<sub>2</sub> by 25 times. Methane  
20  
21 emissions are significant, account for 11% of overall GHG emissions (CO<sub>2</sub> eq.) in US, and in which  
22  
23 landfills account for 20% and natural gas and petrol accounts for 33% of methane emissions <sup>6</sup>. From  
24  
25 these statistics, landfilling of MSW represents one of the highest methane emission sources.  
26  
27  
28  
29

### 30 31 ***Manure generation and management***

32  
33 According to the EPA 2012 report, 20% of the world's non-CO<sub>2</sub> GHG emissions are created  
34  
35 from animal agriculture, and in the US, agriculture GHG emissions account for 9% of the total  
36  
37 emissions, which is 618 million metric tons out of 6870 total million metric tons of CO<sub>2</sub> equivalents <sup>6</sup>.  
38  
39 Animal agriculture emissions include mostly enteric fermentation, the respiration of cattle and other  
40  
41 animals. Manure management has great environmental impact, as it accounts for 14% of the overall  
42  
43 GHG emissions from the agriculture sector. From among the US overall manure management  
44  
45 emissions, nearly 43% of CH<sub>4</sub> emissions are from dairy farms <sup>7</sup>. Key sources for manure production in  
46  
47 the US are cattle, swine and poultry, among which the cattle produced 920 million wet tons of manure,  
48  
49 poultry 80 million wet tons, which includes litter, and swine production accounts for 110 million wet  
50  
51 tons of manure in the year 2007 <sup>8</sup>. By considering the percentages of nutrient concentrations from  
52  
53  
54  
55  
56  
57  
58  
59  
60

1  
2 Combs et al. 1998, it is estimated to represent 7.44 million tons of N and 2.58 million tons of P from  
3  
4 total manure (cattle, swine and poultry), assuming the moisture contents of cattle, swine, and poultry are  
5  
6 76, 80 and 40 respectively <sup>9</sup>. In comparison, US annual agricultural field application of commercial  
7  
8 fertilizers in 2007 is about 13.1 million equivalent tons of N and 4.5 million equivalent tons of P. The  
9  
10 most conventional manure management systems include: 1. Uncovered anaerobic lagoons, 2. Digesters  
11  
12 (includes covered anaerobic lagoons), 3. Solid manure storage, 4. Dry Lots (includes feedlots), 5.  
13  
14 Storage pits, 6. Liquid or Slurry systems, 7. Deep bedding systems (cattle and swine), 8. High-rise  
15  
16 houses for poultry production without litter, 9. Poultry production with litter, 10. Aerobic treatment, and  
17  
18 11. Manure composting <sup>10</sup>. From the Intergovernmental Panel on Climate Change <sup>11</sup> manure methane  
19  
20 emission factors are from 0.02 (most of the poultry breeds), to 1 (beef cattle) and 53 (dairy cows)  
21  
22 kilograms per head per year <sup>12</sup>, US EPA estimates the total methane emissions of 2.478 million tons of  
23  
24 CH<sub>4</sub>/yr. from livestock <sup>12</sup>.

### 30 ***Composting for management of food waste***

31  
32 According to US 2014 statistics, there were 347 composting facilities accepting food waste from  
33  
34 36 states in the US, with 87 accepting mixed organics (leaves, vegetable scraps, tea bags etc.).  
35  
36 Composting consumes more energy than land filling, but significant energy savings in composting are  
37  
38 due to the compost replacement of chemical fertilizer in agriculture <sup>13</sup>. Compost, when applied to the  
39  
40 field, has benefits such as reducing water runoff, soil erosion, and enhancing the metabolism of  
41  
42 microorganisms, which improves the soil fertility. On the other side, it also has a negative impact on the  
43  
44 environment such as CH<sub>4</sub> and CO<sub>2</sub> emissions from compost piles, and uses fossil fuel for transportation  
45  
46 and in the composting equipment. Out of 254 million tons of MSW in 2013, only 3% of the 37 million  
47  
48 tons of food waste is diverted from landfills to composting and it is reported that the composting  
49  
50 methane emissions in the US are 3.3 million tons of CO<sub>2</sub> equivalents <sup>5</sup>.

1 Both landfilling and composting of food waste has a high potential for uncontrolled methane  
2 emissions, so there is a necessity for reliable alternatives for the management of food waste that is  
3 produced each year. Biological treatments are the alternate way for the reduction of solid waste residues  
4 by biological activity<sup>14-17</sup>. Anaerobic treatment of food waste and manure has the potential to reduce  
5 methane emissions from both controlled and uncontrolled sources, as our study shows.  
6  
7  
8  
9  
10  
11  
12

### 13 **LCA Literature review**

14  
15  
16 LCA studies have found AD to be more advantageous (i.e. it has less environmental impacts)  
17 than other organic waste disposal methods, such as incineration and landfilling. Most of the prior LCA  
18 work reported lower greenhouse gas emissions for Bio-CH<sub>4</sub> as a transportation fuel from different  
19 substrates such as energy crops and food waste compared with the fossil methane<sup>18-20</sup>. Betzabet et al  
20 conducted an LCA on optimization of AD process to see the effect of adsorption and desorption in the  
21 process<sup>21</sup>. Owen et al. performed an LCA on the emissions from manure management processes and  
22 their studies highlighted the areas to concentrate on to mitigate the GHG emissions<sup>7</sup>. ROU conducted an  
23 LCA of a windrow composting system in Australia including compost use and post application impacts  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

Only a few LCA studies have analyzed transient landfill emission scenarios, and results show that the  
gas collection systems with flaring had higher emissions when compared with a landfill gas-to-energy  
scenario<sup>23</sup>. A prior life cycle carbon footprint found that emissions were much lower for windrow  
composting, high solids anaerobic digestion, or for co-digestion of the organic fraction of MSW with  
either industrial wastes or sewage sludge compared with the baseline process of composting of the food  
waste while landfilling of the remaining organic fraction<sup>24</sup> (with electricity generation of captured  
landfill gas). However, to date as far as we know no studies have analyzed the entire consequential life  
cycle assessment of Bio-CH<sub>4</sub> produced from AD of mixtures of food waste and dairy manure.

## Research Objectives

The main research objective of this work is to model the cradle-to-grave environmental impacts (greenhouse gas emissions and fossil energy demand) of the anaerobic digestion of food waste mixed with dairy manure at a specific location near Denver, CO. The prior practice at this location was composting of the food waste and manure, and this prior case is modeled as a comparison. Additional analyses include a more general case of diverting food waste from landfills and animal manure from conventional manure management processes to produce Bio-CH<sub>4</sub> in AD facilities. Our study also investigates avoiding different landfilling scenarios (uncontrolled, gas collection and flaring and gas collection and electricity generation operating at steady-state). Finally, in an effort add more realism to the LCA modeling an investigation of the transient response of avoided landfill emissions was conducted.

## Materials and Methods

### *Goal and Scope*

The goal of this study is to make a consequential comparative life cycle analysis in two separate cases. Case 1 analyzes the use of food waste from Denver, CO restaurants and dairy manure in the vicinity of LaSalle, CO for compost production, which was the prior use of these waste materials, versus a new Bio-CH<sub>4</sub> production system. Case 2 is a more general case for the prior use of food waste and dairy manure that assumes food waste was landfilled with (Scenario-1) uncontrolled emissions from the landfill, (Scenario-2) a gas collection system with flaring of the collected landfill gases, and (Scenario-3) which represents landfilling of the food waste with a gas collection system and electricity generation. In all scenarios, conventional manure management is included as part of the avoided pathways. The impact category of primary interest is greenhouse gas emissions; however, fossil energy consumption is also

1  
2 evaluated. Scenario analyzes are modified to include transient landfill emissions that are avoided when  
3  
4 food waste is instead used for Bio-CH<sub>4</sub> production.  
5

### 6 7 *Functional Unit*

8  
9 The basis for the analyses reported here in both cases is the processing of both the food waste  
10 and dairy manure for one day of operation of the Heartland AD facility near LaSalle, CO. This basis  
11 translated to 216 Mg (dry) of food waste and 62 Mg (dry) dairy manure converted in the AD process.  
12  
13 The functional units we have chosen are mainly used in the analyses as a reference to compare the LCA  
14 results to the different alternatives. In Case 1 the functional unit is on the basis of 1 day of operation of  
15 the Heartland AD Bio-CH<sub>4</sub> facility to compare to the business-as-usual (BAU) composting of these  
16 wastes, and Case 2 is on the basis of 1 kg of Bio-CH<sub>4</sub> produced and used in place of BAU landfilling.  
17  
18 All the calculations are based on the total feedstock input of 278 Mg (dry basis) / day which results in an  
19 output Bio-CH<sub>4</sub> of 99.15 Mg, which constitutes 35.66% of initial feedstock (food waste + manure).  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29

### 30 31 *System boundary*

32  
33 In Case 1, the framework for the analysis is designed to model the change to the environment  
34 when the pathways in the BAU composting system are changed to the Bio-CH<sub>4</sub> system. Both systems  
35 must provide the same societal benefit for production of compost, nutrients for agriculture, and energy  
36 from methane (or fossil natural gas). The system boundary for Case 1 BAU composting system is shown  
37 in Figure 1 and includes different pathways that are operating in the prior use of the food waste and  
38 manure. Pathway 1 is composting of food waste and manure. The composting process accounts for the  
39 emissions from transportation of feedstock (food waste, manure and wood chips), windrow process  
40 emissions (including decomposition), and compost land application emissions. The emissions from the  
41 composting process are calculated from prior studies<sup>21,25,26</sup> and are presented in the section-1 of the  
42 Supporting Information document. Pathway 2 includes manufacturing and use of synthetic fertilizers  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55



1  
2 that will be replaced by nutrients in digestate that is produced in the AD system. This accounts for the  
3  
4 emissions from manufacturing, transportation to fields and field application (calculation for synthetic  
5  
6 fertilizer emissions from cradle-to-grave are considered from US average fertilizer mix attested to in  
7  
8 section-1 of Supporting Information document). Pathway 3 in the BAU case is the fossil natural gas that  
9  
10 is replaced by the Bio-CH<sub>4</sub> produced in the AD system, and therefore fossil natural gas is modeled.  
11  
12 Accounted for were all the emissions from cradle-to-grave of fossil natural gas (extraction, process,  
13  
14 transport, and combustion).  
15  
16

17  
18 The Case 1 Bio-CH<sub>4</sub> system shown in Figure 2 also includes multiple pathways and processes.  
19  
20 The first is AD of food waste and manure including the whole lifecycle of the AD Bio-CH<sub>4</sub> pathway  
21  
22 starting from the transportation of food waste and manure until the end use of Bio-CH<sub>4</sub> product,  
23  
24 byproduct compost and digestate (input data provided by Ag Energy). Secondly, compost pathway  
25  
26 produces more compost from food waste than the AD Bio-CH<sub>4</sub> pathway, therefore to provide the same  
27  
28 societal benefit, the remaining compost is made up by equivalent amounts of peat imported from  
29  
30 Canada. The peat pathway considers all emissions from cradle-to-grave. The data regarding the  
31  
32 infrastructure for storage of food waste and manure, for biogas production and purification are not  
33  
34 included in the study because their impacts are assumed negligible for facilities that last decades when  
35  
36 compared to the material and energy inputs to the processes during the life of the AD facility.  
37  
38  
39  
40  
41  
42  
43

44 Figure 1: System boundaries for the Case 1 Business-as-Usual (BAU) system

45  
46 Figure 2: System boundaries for the Case 1 Bio-CH<sub>4</sub> system  
47  
48  
49  
50

51 In Case 2, a modeling approach equivalent to that in Case 1 has been taken. However, LCA  
52  
53 results can be obtained by taking the difference between two systems (BAU landfill system and Bio-CH<sub>4</sub>  
54  
55

1 system) and combining into a single Bio-CH<sub>4</sub> system by considering the credit for emissions from  
2 avoided BAU landfill systems. This approach is a consequential LCA because, in addition to modeling  
3 the direct AD Bio-CH<sub>4</sub> pathway, avoided emissions are also modeled for all co-products that displace  
4 others in the market as well as avoided pathways in the BAU of landfilling and manure management.  
5  
6 Three scenario analyses are studied in case 2 illustrated in Figures 3 & 4. The system boundary for case  
7 2, scenario-1 (Bio-CH<sub>4</sub> system with avoiding uncontrolled landfilling) has multiple pathways in its life  
8 cycle: 1. AD Bio-CH<sub>4</sub> pathway, 2. avoiding uncontrolled landfilling, 3. avoiding manure management, 4.  
9 avoiding synthetic fertilizers and 5. avoiding peat production and use. Avoided uncontrolled landfill  
10 pathway includes the emissions from the transportation of food waste from Denver to the landfill and  
11 emissions during the landfill process (see Supporting Information section 3, Tables 53, 56-58 for  
12 details). Bio-CH<sub>4</sub> system uses manure for production of Bio-CH<sub>4</sub>, so an emissions credit is taken  
13 assuming an equal amount of manure is managed by anaerobic lagoons. The emissions from manure  
14 management are calculated from Silver et al 2015<sup>7</sup>. The compost produced in the AD Bio-CH<sub>4</sub> pathway  
15 is assumed to be met by an equivalent amount of Canadian peat in the BAU system so the emissions  
16 from the peat pathway are considered as credit to the Bio-CH<sub>4</sub> system. Digestate produced in the Bio-  
17 CH<sub>4</sub> system is assumed to displace the manufacture and application of synthetic fertilizers, so the  
18 emissions from synthetic fertilizers are accounted for as a credit to the Bio-CH<sub>4</sub> system and the emission  
19 factors for fertilizer land application as well as compost land application are taken from EPA's Waste  
20 Reduction Model (WARM)<sup>27</sup>.

21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49 Figure 3: System boundaries for the Case 2 Bio-CH<sub>4</sub> scenario avoided BAU with different landfill  
50 systems

51  
52  
53  
54 Figure 4: Case 2 Bio-CH<sub>4</sub> avoided BAU with different landfill scenarios pathways

1  
2  
3  
4 Case 2, Scenario-2 Bio-CH<sub>4</sub> system with avoided landfill gas collection and flaring includes  
5  
6 multiple pathways, which are the same as Scenario 1 except for the avoided pathway of landfilling with  
7  
8 gas collection system and flaring (see Supporting Information section 3, Tables 54, 56-58 for details).  
9  
10 Case 2, scenario-3 Bio-CH<sub>4</sub> system with avoided landfill gas collection and electricity generation also  
11  
12 includes multiple pathways like scenario-2, except the avoided landfill electricity is made up by the  
13  
14 electricity generated with natural gas (see Supporting Information section 3, Tables 55,56-58 for  
15  
16 details).  
17  
18  
19

### 20 ***Allocation***

21  
22 Case 1 is a consequential system, and therefore no allocation is needed. For Case 2, allocation is  
23  
24 avoided by expanding the system boundary to include environmental savings from displacing production  
25  
26 of materials in the market by co-products from Bio-CH<sub>4</sub> production (fertilizers and peat) and by  
27  
28 avoiding the food waste landfilling and manure anaerobic lagoon storage for the waste management.  
29  
30 Peat is selected as an alternative for compost in the Bio-CH<sub>4</sub> life cycle in both cases. Both compost and  
31  
32 peat have different characteristics (Compost consists of humic carbon and peat does not) but can be  
33  
34 compared on a 1:1 volume basis<sup>25</sup>. AD digestate fertilizers replace synthetic fertilizers in the analysis.  
35  
36  
37  
38

### 39 ***Inventory Analysis***

40  
41 All the inputs to the different pathways in Cases 1 and 2 are listed in Tables 1-11 below (the  
42  
43 reader is referred to section-3 in SI document for more information about specific eco-profiles used).  
44  
45 Different eco-profiles are used to calculate the impacts of each process input in the pathways (section-1  
46  
47 in SI document). Emissions in CO<sub>2</sub> equivalents for peat manufacturing, packaging, transport and market  
48  
49 for both scenarios are from a Canadian peatmoss study<sup>28</sup> listed in Table 12.  
50  
51  
52

### 53 ***Impact Assessment***

1  
2 The main impacts analyzed in this study are: 1) global warming and 2) fossil energy use. Energy  
3 use impacts are quantified using the Cumulative Energy Demand method in SimaPro version 8.0.3.14,  
4 which accounts for all the process energy conversion efficiencies and other energy uses from cradle-to-  
5 grave. The global warming impacts are calculated by the IPCC 2013 GWP 100, a method in SimaPro  
6 with global warming potentials (GWP) of CO<sub>2</sub>: 1, N<sub>2</sub>O: 265, CH<sub>4,fossil</sub>: 28 and CH<sub>4,biogenic</sub>: 25.25. GWP  
7 values for all other greenhouse gases included in the developed inventory, such as refrigerants and some  
8 solvents, are also included in this method.  
9  
10  
11  
12  
13  
14  
15  
16  
17

## 18 **Results and Discussion**

### 19 *Case 1 BAU composting versus Bio-CH<sub>4</sub>*

20  
21 The GHG emissions from Case 1 are shown in Tables 13 and 14. The BAU composting system  
22 has total net emissions of  $9.7 \times 10^5$  kg CO<sub>2</sub> eq. /day for 216 tons/day of dry food waste feedstock and 62  
23 tons/day dry dairy manure. Main emission sources include, in decreasing order, composting process  
24 emissions, natural gas production and emissions, and field emissions for compost application. The Bio-  
25 CH<sub>4</sub> system emissions shown in Table 14 total  $8.2 \times 10^5$  kg CO<sub>2</sub> eq./day for the same input rate of dry  
26 food waste and dairy manure and is mainly caused by Bio-CH<sub>4</sub> combustion, AD process-CO<sub>2</sub> separation,  
27 and peat manufacture-use. From this work, the Bio-CH<sub>4</sub> system exhibits 15.5% lower emissions than  
28 the BAU composting system.  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41

42 For fossil energy use in case 1, the Bio-CH<sub>4</sub> system is calculated based on the input energy  
43 supply by electricity and natural gas for the AD process, transportation, peat process, and output energy  
44 production from Bio-CH<sub>4</sub> system with the help of SimaPro Cumulative Energy Demand method. Fossil  
45 energy use for Bio-CH<sub>4</sub> system is calculated to be  $2.93 \times 10^6$  MJ/day, where as in the BAU composting  
46 system energy use was calculated to be  $6.3 \times 10^6$  MJ/day, which includes the energy usage for process  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

1  
2 operations in natural gas, composting, and fertilizer pathways. When comparing the Bio-CH<sub>4</sub> system  
3  
4 with the BAU composting system, the Bio-CH<sub>4</sub> system has 53.4% savings as shown in the Figure 5.  
5  
6  
7

8  
9 Figure 5: Fossil energy demand for BAU composting system and Bio-CH<sub>4</sub> system in units of MJ/day  
10  
11  
12  
13  
14  
15

### 16 *Case 2 Scenario steady state analyses*

17  
18 The GHG emissions from individual pathways in scenario-1, scenario-2 and scenario-3 are  
19 shown in Figure 6. All emission calculations for CO<sub>2</sub> and CH<sub>4</sub> emissions in the three scenarios are  
20 reported in the Supporting Information document section 2. In the steady-state scenario-1, the Bio-CH<sub>4</sub>  
21 system with the avoided BAU uncontrolled landfill system has -17.76 kg CO<sub>2</sub> eq. / kg of Bio-CH<sub>4</sub>.  
22 Scenario-2 with avoided BAU landfill with gas collection system and flaring has -5.49 kg CO<sub>2</sub> eq. / kg  
23 of Bio-CH<sub>4</sub> production. Scenario-3 with avoided BAU landfill with gas collection and electricity  
24 generation has -3.51 kg CO<sub>2</sub> eq. / kg of Bio-CH<sub>4</sub> production. The main advantage of the AD process for  
25 production of Bio-CH<sub>4</sub> is the avoiding of landfill and manure management emissions because the  
26 savings are much greater than the emissions from the AD Bio-CH<sub>4</sub> pathway. Emissions from the AD  
27 Bio-CH<sub>4</sub> pathway are 6.1 kg CO<sub>2</sub> eq. / kg Bio-CH<sub>4</sub>, without considering the substantial avoided  
28 emissions, and by its own is higher than fossil natural gas (4.3 kg CO<sub>2</sub> eq. / kg Bio-CH<sub>4</sub> eq. of fossil  
29 natural gas). In all scenarios, savings of emissions are greatest for the avoided landfill pathway, followed  
30 by avoided manure management, then peat production, and finally avoided synthetic fertilizers.  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

Figure 6: Case 2, Bio-CH<sub>4</sub> emissions avoiding different BAU steady-state landfill scenarios

### Case 2 transient analysis

When food waste is deposited in a landfill, emissions of CH<sub>4</sub> and CO<sub>2</sub> are not generated to full potential immediately, but instead the landfill AD processes require several decades to complete the biomass decomposition. In order to introduce a more accurate avoided landfill emissions calculation, a transient model for CH<sub>4</sub> and CO<sub>2</sub> generation and emissions from landfills was derived in section 2 of the SI document for both uncontrolled landfills and for landfills with gas collection and flaring systems. The equation representing the avoided landfill emission rate for both CH<sub>4</sub> and CO<sub>2</sub>, respectively, are

$$= M/2(1 - e^{-kt}) * 16/12$$

$$= M/2(1 - e^{-kt}) * 44/12$$

, where  $M$  is the annual rate of food waste landfilled (metric tons carbon/yr.), which in our calculations is the same as the annual food waste C input rate to the AD Bio-CH<sub>4</sub> facility,  $k$  is a first order reaction rate constant for decomposition of food waste and production of Bio-CH<sub>4</sub> and CO<sub>2</sub> by the AD process (0.12 yr<sup>-1</sup>), and  $t$  is time in years<sup>26</sup>. These avoided landfill emissions increase exponentially until steady-state is achieved after approximately 50 years (see section 2 of the SI for derivations).

Case 2 transient results are modeled for three scenarios. In Scenario-1, the Bio-CH<sub>4</sub> system avoids the BAU landfill assuming uncontrolled emissions. In scenario-2, the Bio-CH<sub>4</sub> system substitutes for a BAU landfill emissions with gas collection system and flaring. In scenario-3, the Bio-CH<sub>4</sub> system substitutes for a BAU landfill emissions with gas collection and electricity generation. Figure 7 shows transient emissions from scenario-1, 2 and 3 over a 50-year time frame based on one kg of generated Bio-CH<sub>4</sub>. The Bio-CH<sub>4</sub> emissions in scenario-1 drop sharply over the 50-year simulation. From our simulation, emissions from Bio-CH<sub>4</sub> are always lower than fossil natural gas over the entire simulation period from years 1 – 50 (except for scenario-1 at 1 year). The Bio-CH<sub>4</sub> system for

1  
2 scenario-1 has a savings of -17.76 kg eq. CO<sub>2</sub> emissions for a kg of Bio-CH<sub>4</sub> by the end of the 50-yr  
3  
4 cycle.  
5  
6  
7

8  
9 Figure 7: Case 2, Bio-CH<sub>4</sub> emissions avoiding the BAU landfilling a) uncontrolled Scenario-1,  
10  
11 b) gas collection system (GCS) & flaring Scenario-2, and c) landfill gas collection with electricity  
12  
13 generation (LFGE) Scenario-3  
14  
15  
16  
17

18  
19 In scenario-2 most of the methane emitted in the uncontrolled landfill scenario is captured  
20  
21 (assumption of 75% collection efficiency and is flared) using a gas collection system and flared as CO<sub>2</sub>  
22  
23 thereby avoiding the emission of high GWP CH<sub>4</sub>. The simulated results of the transient scenario-2 are  
24  
25 shown in the Figure 7. In scenario-2, GHG emissions of AD Bio-CH<sub>4</sub> pathway are always lower than  
26  
27 fossil natural gas over the entire 50-year modeling time. After 50 years avoiding the emissions from  
28  
29 BAU with landfill gas collection and flaring system, the Bio-CH<sub>4</sub> emits -5.49 kg CO<sub>2</sub> eq./kg Bio-CH<sub>4</sub>,  
30  
31 thus reducing global warming effect with each unit of production of Bio-CH<sub>4</sub>. These favorable  
32  
33 emissions are compared to that of fossil natural gas, 4.3 kg CO<sub>2</sub> eq. / 1.06 kg fossil natural gas, which  
34  
35 contains mostly fossil methane [30], where the factor of 1.06 is the ratio of LHV of BioCH<sub>4</sub> (50 MJ/kg)  
36  
37 to fossil natural gas (47.14 MJ/kg). This comparison shows the significant benefit of the AD Bio-CH<sub>4</sub>  
38  
39 pathway with respect to fossil natural gas. Scenario-3 is also shown in Figure 7 and has an overall  
40  
41 savings of -3.51 kg CO<sub>2</sub> eq. /kg Bio-CH<sub>4</sub> by the end of 50 years.  
42  
43  
44  
45

#### 46 ***Fossil Energy Consumption for Case 2***

47  
48  
49 The steady-state fossil energy consumption for the Bio-CH<sub>4</sub> system when avoiding the BAU  
50  
51 landfill, either uncontrolled or gas collection and flaring, is calculated as  $1.69 \times 10^6$  MJ/99.15 tons of Bio-  
52  
53 CH<sub>4</sub>, or 17.06 MJ/kg Bio-CH<sub>4</sub>. By comparison, fossil natural gas uses 59.0 MJ fossil energy /1.06 kg  
54  
55

1 fossil natural gas. Fossil energy demand for the Bio-CH<sub>4</sub> system scenario-3 is calculated as 49.26 MJ/kg  
2 Bio-CH<sub>4</sub>. Bio-CH<sub>4</sub> fossil energy consumption is 71.1 % lower than fossil natural gas in scenario-1 and  
3  
4 Bio-CH<sub>4</sub>. Bio-CH<sub>4</sub> fossil energy consumption is 71.1 % lower than fossil natural gas in scenario-1 and  
5  
6 scenario-2, and is 16.5 % lower in scenario-3. Figure 8 shows the contributions by different processes in  
7  
8 the AD Bio-CH<sub>4</sub> toward the cumulative fossil energy consumption. The AD process is the largest  
9  
10 contributor to fossil energy demand among all of the pathway stages in scenario-1 and scenario-2. The  
11  
12 makeup of avoided landfill electricity generation using natural gas is the largest contributor to fossil  
13  
14 energy demand in scenario-3.  
15  
16  
17  
18  
19  
20

21 Figure 8: Fossil energy demand for case 2, scenario-1, scenario-2, and scenario-3  
22  
23  
24

25 By summarizing the discussed GHG emission results from our study, we show that AD of food waste  
26  
27 and dairy manure in scenarios that avoid landfilling provides the best reductions in GHG emissions  
28  
29 compared to composting of the wastes. A similar conclusion was made from the study by Yoshida et al.  
30  
31 on AD of food waste and organics in MSW<sup>23</sup>, which also showed the best GHG emission savings for  
32  
33 avoided landfilling rather than composting.  
34  
35  
36

### 37 **Broader impacts of AD in the US**

38  
39 If implemented on a national scale in the US, a savings of 0.74% of the present annual natural  
40  
41 gas energy demand can be realized from production of Bio-CH<sub>4</sub> through AD of all food waste and a  
42  
43 significant fraction of total dairy manure. Our results show that per kg of Bio-CH<sub>4</sub> produced from food  
44  
45 waste and dairy manure, net emissions after avoiding landfilling with gas collection and flaring and  
46  
47 avoiding conventional dairy manure management saves 5.49 kg CO<sub>2</sub> eq. emissions, the uncontrolled  
48  
49 landfilling scenario saves 17.76 kg CO<sub>2</sub> eq., and the gas collection and electricity generation landfilling  
50  
51 scenario saves 3.51 kg CO<sub>2</sub> eq. These savings do not factor in the additional emissions savings when  
52  
53  
54  
55  
56



1 fossil natural gas is displaced, so net savings will be larger still. From data on landfilling there are 850  
2  
3 landfills with flaring and gas collection, 400 uncontrolled landfills, and 658 landfills with gas collection  
4  
5 and electricity generation, which if avoided in the future through AD of food waste would provide a  
6  
7 weighted average savings of 7.37 kg eq. CO<sub>2</sub> emissions per kg of Bio-CH<sub>4</sub> produced from food waste  
8  
9 and manure blend.  
10  
11

12  
13 Based on the ratios provided by AgEnergy, 19 million short tons of manure from dairy  
14  
15 production in the US alone can provide sufficient blending for the total 36 million tons of food waste  
16  
17 that is landfilled. By diverting the food waste and manure to anaerobic digestion, approximately 0.41%  
18  
19 of overall GHG emissions of the approximately seven billion tons CO<sub>2</sub> eq. can be saved in the US  
20  
21 annually using approximately 100 Heartland-scale AD facilities. It is important to point out other  
22  
23 sustainability benefits of Bio-CH<sub>4</sub> production from food waste and dairy manure beyond conservation of  
24  
25 fossil energy and reductions in GHG emission. These potential benefits include the recycling of mineral  
26  
27 nutrients from food waste and manure to agricultural fields and associated reduction of environmental  
28  
29 impacts of in synthetic fertilizer production and conserving natural resources. Although water is  
30  
31 consumed in the AD process, digestate water is delivered to surrounding agricultural fields to offset  
32  
33 some irrigation water usage. Large-scale deployment of AD of food waste / manure mixtures in the US  
34  
35 would stimulate economic growth and create many engineering, facility operator, and spinoff jobs.  
36  
37 More comprehensive sustainability analyses should be conducted to better understand the full set of  
38  
39 potential benefits and costs (loss of jobs in landfilling and natural gas industries, possible odor issues)  
40  
41 from large-scale production of Bio-CH<sub>4</sub> in the US.  
42  
43  
44  
45  
46  
47  
48

### 49 **Conclusions**

50  
51 This paper investigated lifecycle GHG emissions of Bio-CH<sub>4</sub> production from food waste and  
52  
53 dairy manure. This study showed that the Bio-CH<sub>4</sub> system emits lower greenhouse gases and requires  
54  
55

1 less net fossil energy during its whole lifecycle when compared to an equivalent amount of fossil natural  
2 gas. It also has lower emissions than other treatment processes in the US for these solid wastes, because  
3  
4  
5  
6 the GHG emissions savings from avoiding the conventional management of manure and landfilling of  
7  
8  
9 food waste have a significant benefit on the overall GHG reduction potential of Bio-CH<sub>4</sub>.

### 10 11 **Acknowledgments**

12  
13 We wish to thank the Richard and Bonnie Robbins Endowment and the Sustainable Futures  
14  
15  
16 Institute at Michigan Technological University for financial support for Sharath K Ankathi in this  
17  
18  
19 research.  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

## References

1. U.S. EPA, Advancing Sustainable Materials Management: 2013 Fact Sheet. June 2015; [www.epa.gov/sites/production/files/2015-09/documents/2013\\_advncng\\_smm\\_fs.pdf](http://www.epa.gov/sites/production/files/2015-09/documents/2013_advncng_smm_fs.pdf).
2. Statista. Distribution of municipal waste disposal and recovery in the United States in 2013 by treatment. 2013; [www.statista.com/statistics/478944/disposal-and-recovery-of-us-municipal-waste-distribution-by-treatment/](http://www.statista.com/statistics/478944/disposal-and-recovery-of-us-municipal-waste-distribution-by-treatment/). Accessed 10/02, 2016.
3. Levis J, Barlaz MA, Themelis NJ, Ulloa P. Assessment of the state of food waste treatment in the United States and Canada. *Waste Management*. 2010;30(8):1486-1494.
4. Powell JT, Townsend TG, Zimmerman JB. Estimates of solid waste disposal rates and reduction targets for landfill gas emissions. *Nature Climate Change*. 2016;6(2):162-165.
5. U.S. EPA, Inventory of U.S. Greenhouse Gas Emissions and Sinks:1999-2012. April 15, 2014:Chapter 8, 8-32.
6. U.S. EPA, Inventory of U.S. Greenhouse Gas Emissions and Sinks: 1999-2014 2016.
7. Owen JJ, Silver WL. Greenhouse gas emissions from dairy manure management: a review of field-based studies. *Global change biology*. 2015;21(2):550-565.
8. He Z, Zhang H. *Applied manure and nutrient chemistry for sustainable agriculture and environment*: Springer; 2014.
9. UWEX A3557. Nutrient Management, Practices for Wisconsin Corn Production and Water Quality Protection. Wisconsin2006.
10. U.S. EPA, Guide for the Agriculture and Livestock Sectors2009:40 CFR part 98.
11. Hongmin Dong JM. Emissions from Livestock and manure management. In: IPCC, ed. *Agriculture, Forestry and Other Land Use*. Vol 42016:87.

- 1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60
12. Hristov AN, Johnson KA, Kebreab E. Livestock methane emissions in the United States. *Proceedings of the National Academy of Sciences of the United States of America*. 2014;111(14).
13. Di Maria F, Micale C. Life cycle analysis of management options for organic waste collected in an urban area. *Environmental Science and Pollution Research*. 2015;22(1):248-263.
14. De Gioannis G, Muntoni A, Cappai G, Milia S. Landfill gas generation after mechanical biological treatment of municipal solid waste. Estimation of gas generation rate constants. *Waste Management*. 2009;29(3):1026-1034.
15. Di Maria F, Sordi A, Micale C. Experimental and life cycle assessment analysis of gas emission from mechanically–biologically pretreated waste in a landfill with energy recovery. *Waste Management*. 2013;33(11):2557-2567.
16. Komilis D, Ham R, Stegmann R. The effect of municipal solid waste pretreatment on landfill behavior: a literature review. *Waste Management and Research*. 1999;17(1):10-19.
17. Fricke K, Santen H, Wallmann R. Comparison of selected aerobic and anaerobic procedures for MSW treatment. *Waste management*. 2005;25(8):799-810.
18. Patterson T, Esteves S, Dinsdale R, Guwy A, Maddy J. Life cycle assessment of biohydrogen and biomethane production and utilisation as a vehicle fuel. *Bioresource technology*. 2013;131:235-245.
19. Langlois J, Sassi JF, Jard G, Steyer JP, Delgenes JP, Hélias A. Life cycle assessment of biomethane from offshore□cultivated seaweed. *Biofuels, Bioproducts and Biorefining*. 2012;6(4):387-404.
20. Stucki M, Jungbluth N, Leuenberger M. Life cycle assessment of biogas production from different substrates. *Bundesamt für Energie BFE Swiss Centre for Life Cycle Inventories (Ecoinvent Centre), ecoinvent-Database*. 2011.

- 1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60
21. Morero B, Groppelli E, Campanella EA. Life cycle assessment of biomethane use in Argentina. *Bioresource technology*. 2015;182:208-216.
  22. Recycled Organics Unit, Life Cycle Inventory and Life Cycle Assessment for Windrow Composting Systems. Department of Environment and Conservation. *New South Wales*. 2007.
  23. Wanichpongpan W, Gheewala SH. Life cycle assessment as a decision support tool for landfill gas-to energy projects. *Journal of Cleaner Production*. 2007;15(18):1819-1826.
  24. Yoshida H, Gable JJ, Park JK. Evaluation of organic waste diversion alternatives for greenhouse gas reduction. *Resources, Conservation and Recycling*. 2012;60:1-9.
  25. Saer A, Lansing S, Davitt NH, Graves RE. Life cycle assessment of a food waste composting system: environmental impact hotspots. *Journal of Cleaner Production*. 2013;52:234-244.
  26. RTI International. *Greenhouse Gas Emissions Estimation Methodologies for Biogenic Emissions from Selected Source Categories: Solid Waste Disposal Wastewater Treatment Ethanol fermentation*: U.S. Environmental Protection Agency; December 14, 2010.
  27. ICF International. *Documentation for Greenhouse Gas Emission and Energy Factors Used in the Waste Reduction Model (WARM)*: U.S. Environmental Protection Agency Office of Resource Conservation and Recovery; February 2016.
  28. Peat and peatland, Canadian environmental life cycle assessment, product fact sheet. 2016; [www.tourbehorticole.com/wp-content/uploads/2015/01/cycle.pdf](http://www.tourbehorticole.com/wp-content/uploads/2015/01/cycle.pdf). Accessed 09/ 06, 2016.

**Table 1:** Inputs to AD Bio-CH<sub>4</sub> pathway (basis of 1 day): Case 1 and Case 2

AD pathway (Bio-CH <sub>4</sub> )	Process	Amount	Unit	
AD pathway (Bio-CH <sub>4</sub> )	1. Manure transportation from local farm to facility	Inbound trip	4.69x10 <sup>3</sup>	t*km
		Return trip	3.75 x10 <sup>3</sup>	t*km
	2. Food waste transportation from Denver to facility	Inbound trip	5.76 x10 <sup>4</sup>	t*km
		Return trip	4.61 x10 <sup>4</sup>	t*km
	3. AD process	Electricity	132	MWh
		Natural gas	3.2 x10 <sup>8</sup>	btu
	4. Bio-CH <sub>4</sub> Combustion emissions (Bio-CH <sub>4</sub> combustion)		99151.2	kg CH <sub>4</sub> combusted
	5. CO <sub>2</sub> Fugitive emissions from AD (CO <sub>2</sub> emissions from AD)		99151.2	kg CO <sub>2</sub> emitted
6. Transportation of compost from AD to Denver market	Inbound trip	9.71 x10 <sup>3</sup>	t*km	
	Return trip	7.77 x10 <sup>3</sup>	t*km	
7. AD digestate N field application (Liquid digestate N <sub>2</sub> O emissions on field application)		2.34x10 <sup>3</sup>	kg of N	
8. AD compost field application (CO <sub>2</sub> emissions from AD compost soil application)		1.22 x10 <sup>5</sup>	kg	

**Table 2:** Inputs to compost pathway (basis of 1 day): Case 1 BAU

Compost pathway (BAU)	Process		Amount	Unit
	1. Manure transportation from local farms to compost facility	Inbound trip		$4.66 \times 10^3$
Return trip			$3.73 \times 10^3$	t*km
2. Wood pallets transportation from nearby	Inbound trip		221.3	t*km
	Return trip		177	t*km
3. Food waste transportation from Denver to compost	Inbound trip		$8.11 \times 10^4$	t*km
	Return trip		$6.49 \times 10^4$	t*km
4. Diesel used in tractor for composting	Tractor		$4.92 \times 10^3$	Kg
	Turner		16.8	kg
	Grinder		594	kg
5. Composting decomposition emissions (CO <sub>2</sub> ) (Composting Decomposition Emissions (CO <sub>2</sub> ))			322.94	tons dry compost
6. Composting decomposition emissions (CH <sub>4</sub> &N <sub>2</sub> O) (Composting Decomposition Emissions (CH <sub>4</sub> and N <sub>2</sub> O))			$1.18 \times 10^3$	tons (wet) compost
7. Compost land application (at 50% moisture) (CO <sub>2</sub> Emissions from Compost Land Application)			395.3	tons wet compost
8. Compost (wet) transportation from compost facility to Denver market	Inbound trip		44550.31	t*km
	Return trip		35640.24	t*km

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

**Table 3:** Inputs to natural gas pathway: Case 1 BAU

Natural gas pathway (BAU)	<b>Process</b>	<b>Amount</b>	<b>Unit</b>
	1.Emissions from natural gas transport, extraction processing, distribution and usage	3.97 x10 <sup>6</sup>	MJ of heat

For Peer Review



**Table 4:** Inputs to synthetic fertilizer pathway: Case 1 BAU

Synthetic fertilizer pathway (BAU)	Process		Amount	Unit
	1. Emissions from synthetic fertilizers manufacturing process and market	Nitrogen		$2.34 \times 10^3$
Phosphate		$0.75 \times 10^3$	kg	
Potassium		$2.91 \times 10^3$	kg	
2. Synthetic fertilizer transport from Denver market to farm	Inbound trip		44.9	t*km
	Return trip		35.9	t*km
3. Fertilizer N field application (Emissions of N <sub>2</sub> O from synthetic N fertilizer applied to Field)			$2.34 \times 10^3$	kg of N

For Peer Review

**Table 5:** Inputs to peat pathway: Case 1 Bio-CH<sub>4</sub>

Peat pathway (Bio-CH <sub>4</sub> )	Process	Amount	Unit	
	1. Emissions from peat moss manufacturing, transport and use (Peat moss Manufacturing, Transport and Use (CO <sub>2</sub> Emissions))	546	m <sup>3</sup> make up peat	
	2. Transportation of Peat moss from Canada(Saskatchewan) to Denver	Inbound trip	1.42 x10 <sup>5</sup>	t*km
		Return trip	1.14 x10 <sup>5</sup>	t*km

For Peer Review

**Table 6:** Inputs to avoided landfill without gas collection pathway: Case 2 Scenario-1

Avoided	<b>Process</b>		<b>Amount</b>	<b>Unit</b>
landfill pathway uncontrolled emissions	1. Transportation of food waste from Denver to landfill	Inbound trip	8.42 x10 <sup>3</sup>	t*km
		Return trip	6.74 x10 <sup>3</sup>	t*km
	2. Emissions from landfill without gas collection CH <sub>4</sub> and CO <sub>2</sub> (Steady state) Uncontrolled		216.6	tons (dry) food waste input to landfill

For Peer Review

**Table 7:** Inputs to avoided landfill with gas collection and flaring: Case 2 Scenario-2

Avoided landfill pathway	<b>Process</b>		<b>Amount</b>	<b>Unit</b>
with gas collection and flaring	1. Transportation of food waste from Denver to landfill	Inbound trip	8.42 x10 <sup>3</sup>	t*km
		Return trip	6.74 x10 <sup>3</sup>	t*km
	2. Emissions from landfill with gas collection and flare CH <sub>4</sub> and CO <sub>2</sub> (Steady state) (Landfill Emissions GCS steady state)		216.6	tons (dry) food waste input to landfill

For Peer Review

**Table 8:** Inputs to avoided landfill with gas collection & electricity pathway: Case 2 Scenario-3

Avoided landfill pathway with gas collection and electricity generation	Process	Amount	Unit	
	1. Transportation of food waste from Denver to landfill	Inbound trip	8.42 x10 <sup>3</sup>	t*km
		Return trip	6.74 x10 <sup>3</sup>	t*km
	2. Emissions from LFGE (landfill with gas collection and electricity generation) CH <sub>4</sub> and CO <sub>2</sub> (steady state)	216.6	tons (dry) food waste input to landfill	
	3. Electricity generation from Natural gas	287	MWh	

For Peer Review

**Table 9:** Inputs to avoided synthetic fertilizer pathway: Case 2

Avoided Synthetic fertilizer pathway	Process		Amount	Unit
	1. Emissions from synthetic fertilizers manufacturing process and market	Nitrogen	$2.34 \times 10^3$	kg
		Phosphate	$0.75 \times 10^3$	kg
		Potassium	$2.91 \times 10^3$	kg
	2. Transportation of synthetic fertilizers from LaSalle market to fields	Onward trip	44.9	t*km
		Return trip	35.9	t*km
	3. Fertilizer N field application (Emissions of N <sub>2</sub> O from synthetic N fertilizer applied to Field)		$2.34 \times 10^3$	kg of N

For Peer Review

**Table 10:** Inputs to avoided peat moss pathway: Case 2

Avoided	<b>Process</b>	<b>Amount</b>	<b>Unit</b>	
peat pathway	1. Emissions from peat moss manufacturing, transport and use (Solid digestate compost from AD) (Peat moss Manufacturing, Transport and Use (CO <sub>2</sub> Emissions))	245	m <sup>3</sup>	
	2. Transportation of Peat moss from Canada(Saskatchewan) to Denver	Inbound trip	6.4 x10 <sup>4</sup>	t*km
		Return trip	5.12 x10 <sup>4</sup>	t*km

For Peer Review

**Table 11:** Inputs to avoided manure pathway: Case 2

Avoided manure pathway	Process	Amount	Unit
	1. Anaerobic lagoon emissions	40.49	tons (dry basis)
	2. Slurry storage tanks	6.23	tons (dry basis)
	3. Solid manure piles	15.58	tons (dry basis)

For Peer Review



**Table 12:** GHG emissions for peat manufacturing, packaging transport and use

Category	Unit	Harvest	Package	Transport	Soil application	In situ decomposition	Total
GHG Emissions	Kg CO <sub>2</sub> equivalent	4.03	2.53	15.63	183	60.79	269.7

For Peer Review

**Table 13** Case 1: Business As Usual composting system (Basis of 1 day)

Process	kg CO <sub>2</sub> eq.	% Contribution
<b>Composting Pathway</b>		
1. Feedstock transport	29,592.63	3.06
2. Composting process equipment fuel use	20,796.49	2.14
3. Decomposition of feedstock in composting process	367,988.1	37.92
4. Transportation from compost facility to market	15,313.93	1.58
5. Land application of compost	173,939.4	17.93
<b>Fossil Natural Gas Pathway</b>		
6. Natural gas production and combustion	317,870.7	32.76
<b>Synthetic Fertilizer Pathway</b>		
7. Synthetic fertilizer production and market	18,904.28	1.95
8. Synthetic fertilizer transport from market to farm	13.92	0.0014
9. Synthetic N fertilizer soil application	25,786.51	2.66
<b>Total</b>	<b>970,206</b>	<b>100</b>

**Table 14:** Case 1: Bio-CH<sub>4</sub> system (Basis of 1 day)

Process	kg CO <sub>2</sub> eq.	% Contribution
<b>AD Bio-CH<sub>4</sub> Pathway</b>		
1. Manure transportation from dairy farm to AD	1,453.99	0.17
2. Food waste transport from Denver to AD facility	17,856.1	2.17
3. Anaerobic digestion Process	134,499.6	16.39
4. CO <sub>2</sub> separation from AD biogas and venting	121,956	14.86
5. Biogas CH <sub>4</sub> combustion	272,665.8	33.23
6. AD digestate soil application	25,786.5	3.14
7. AD compost transport to Denver	3,010.29	0.36
8. AD compost applied to land	53,824.3	6.56
<b>Peat Pathway</b>		
9. Peat manufacturing, transport, use	145,217.1	17.69
10. Peat transport from Saskatchewan to Denver	44,278.6	5.39
<b>Total</b>	<b>820,548.3</b>	<b>100</b>

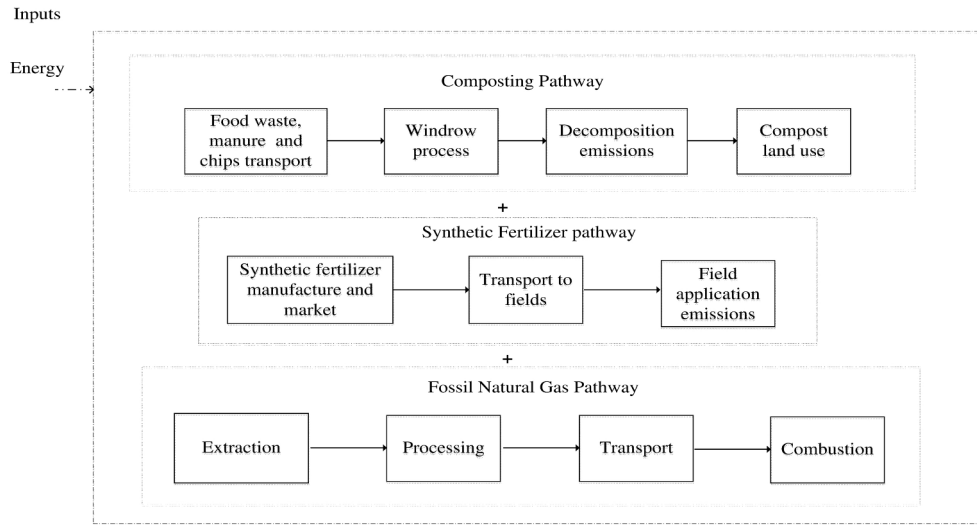


Figure 1: System boundaries for the Case 1 Business-as-Usual (BAU) system

115x62mm (600 x 600 DPI)

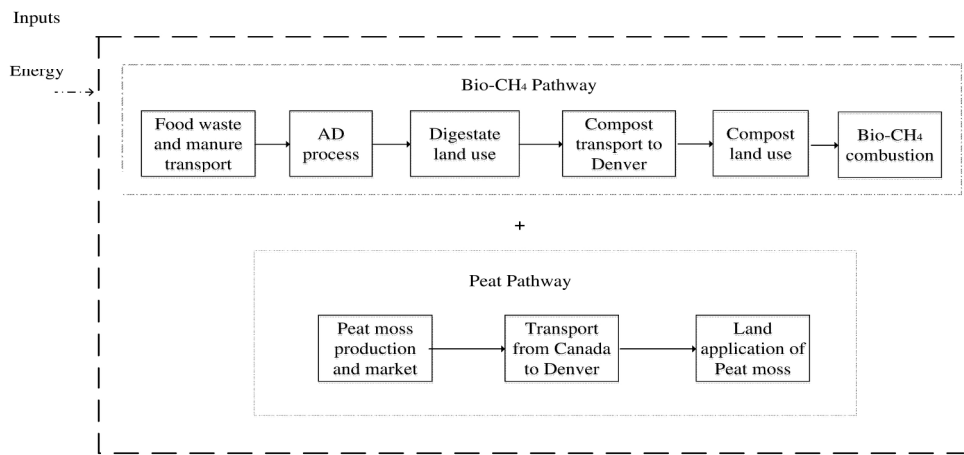


Figure 2: System boundaries for the Case 1 Bio-CH<sub>4</sub> system

111x57mm (600 x 600 DPI)

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

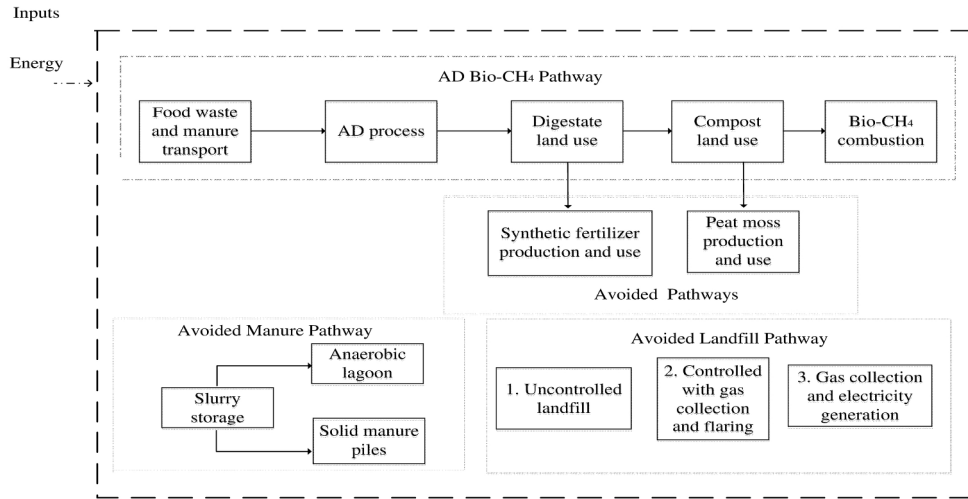


Figure 3: System boundaries for the Case 2 Bio-CH<sub>4</sub> scenario avoided BAU with different landfill systems

112x59mm (600 x 600 DPI)

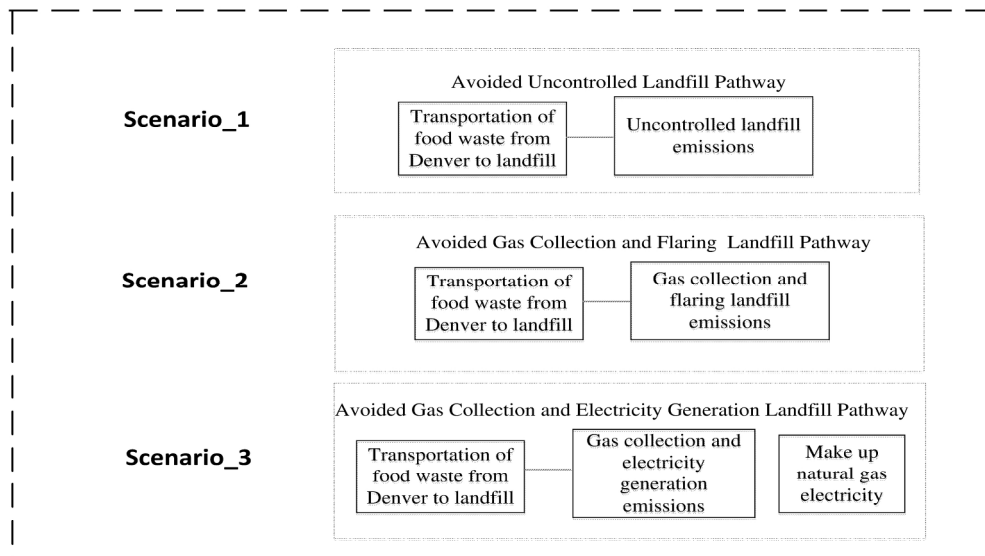


Figure 4: Case 2 Bio-CH4 avoided BAU with different landfill scenarios pathways

111x62mm (600 x 600 DPI)

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

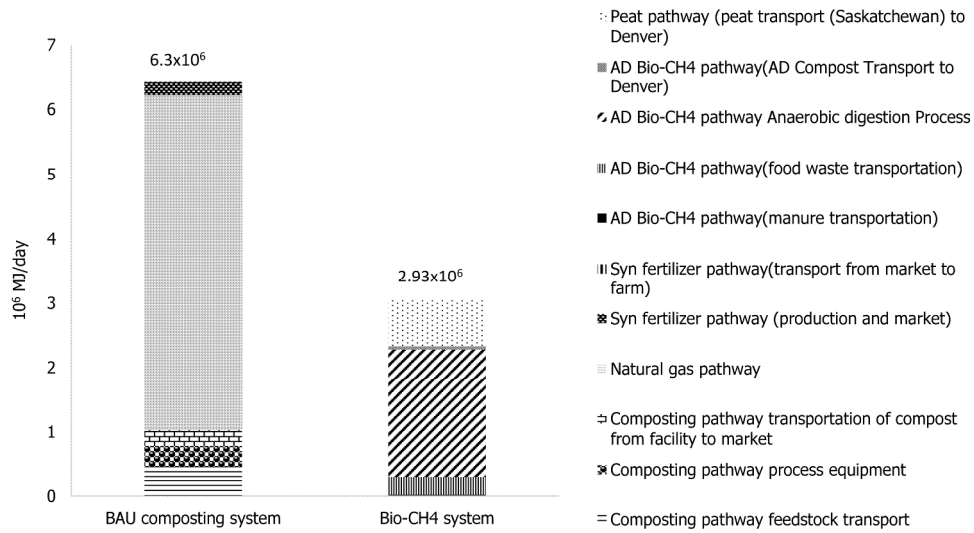


Figure 5: Fossil energy demand for BAU composting system and Bio-CH4 system in units of MJ/day

129x78mm (600 x 600 DPI)



1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

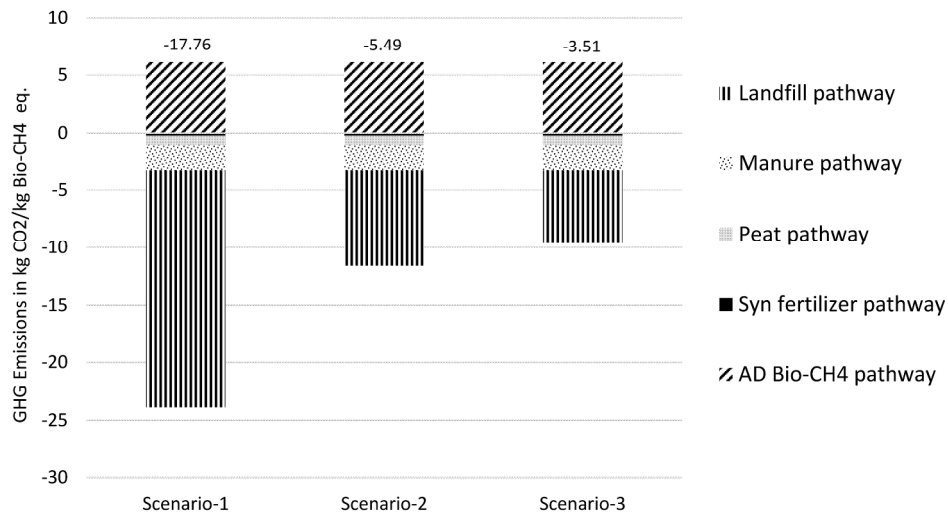


Figure 6: Case 2, Bio-CH<sub>4</sub> emissions avoiding different BAU steady-state landfill scenarios

122x70mm (600 x 600 DPI)

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

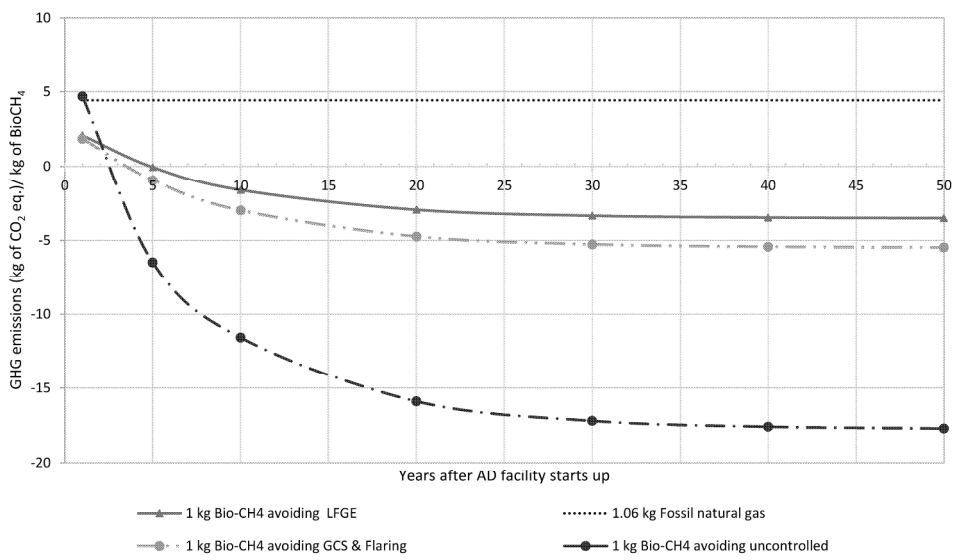


Figure 7: Case 2, Bio-CH4 emissions avoiding the BAU landfilling a) uncontrolled Scenario-1, b) gas collection system (GCS) & flaring Scenario-2, and c) landfill gas collection with electricity generation (LFGE) Scenario-3

121x70mm (600 x 600 DPI)

Review

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

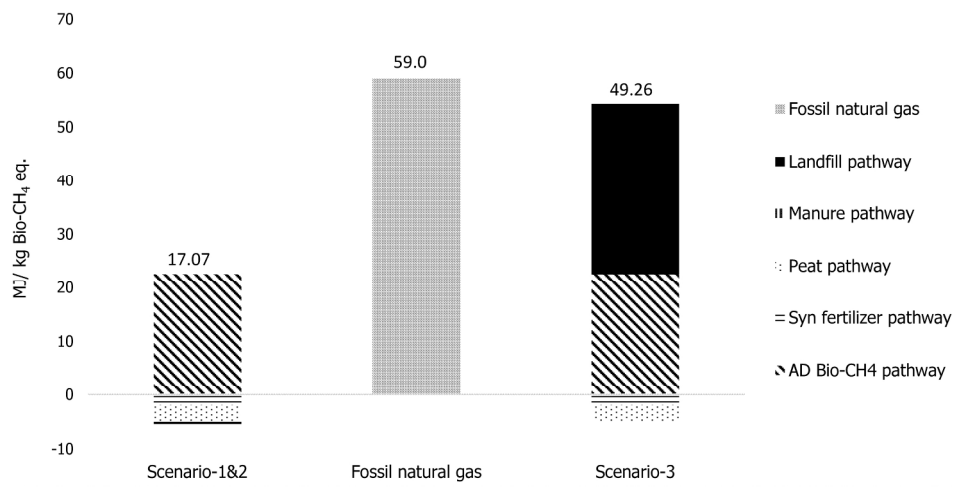


Figure 8: Fossil energy demand for case 2, scenario-1, scenario-2, and scenario-3

108x57mm (600 x 600 DPI)